

# Growth and Physiological Responses of the Singapore Strain of *Heterosigma* to Various Nitrogen Sources and Light Conditions

Jerome Wai Kit Kok

Tropical Marine Science Institute (TMSI)  
National University of Singapore (NUS)  
Singapore, Singapore  
jeromekok@nus.edu.sg

Sandric Chee Yew Leong

Tropical Marine Science Institute (TMSI)  
National University of Singapore (NUS)  
Singapore, Singapore

**Abstract—** *Nutrient pollution of coastal environments particularly by nitrogen (N) sources can lead to harmful algal blooms. Ameliorating the incidence of such events requires an understanding of how harmful taxa respond to different N sources. In particular, the blooms of toxic Heterosigma have been known to cause costly and highly detrimental effects to surrounding ecosystem function. The threat is especially dire in the case of Singapore where growth dynamics of Heterosigma species under tropical environmental conditions are poorly understood. The present study examined how the growth rates of Heterosigma species isolated from the Johor Straits of Singapore differ under various N sources as well as varying light intensity. Heterosigma was found to grow well under both nitrogen sources, but grew at a faster rate when exposed to nitrate compared to ammonium. In addition, the study also observed tolerance for high- and low-light conditions. These results indicate that present modifications to the coastal environment through high sedimentation and nutrient pollution rates might be creating conditions conducive to the growth of Heterosigma.*

**Keywords -** *Heterosigma; Singapore; coastal management; pollution; harmful; bloom*

## I. INTRODUCTION

HABs of the marine raphidophyte *Heterosigma* have been known for massive fish- and shellfish-kills throughout the world [1-3]. *Heterosigma* can cause detrimental effects through neurotoxins [4-7] even without attaining high cell concentration and dominance of the water column [8-9]. In recent years, raphidophyte *Heterosigma* species found in water samples collected around Singapore indicated that they could be dominant at times [10-11]. However, there is at present little knowledge about tropical strains [12], making it difficult to mitigate growth and productivity levels.

Current measures to address the occurrence of *Heterosigma* must seek to ameliorate the formation of HABs which would severely damage the coastal environment. Information on their growth and physiological responses to

varying environmental conditions are therefore essential for implementing measures that can protect commercial aquaculture as well as public health.

It is widely recognised that marine environments are N-limiting for marine algae growth [13]. In general, increased agricultural development as well as sewage and industrial output have led to increased inorganic-N levels in coastal environments [14] which are commonly in the forms of ammonium and nitrate. In Singapore, its waters are similarly susceptible to N-pollution due to nearby coastal and urban development [15]. N-pollution can enter marine environments through surface run-off, atmospheric deposition or groundwater discharge [16], entering coastal waters at different depths and resulting in vertical inhomogeneity of inorganic-N levels.

In the present study, we therefore exposed *Heterosigma* species to different light exposure and different form of N-sources and concentrations to examine their growth responses to such varying conditions.

## II. MATERIALS AND METHODS

*Heterosigma* strains used in the experiment were isolated from the East Johor Straits by Dr. Sandric Leong. Cultures grown in 1-litre sterilised screw-top polycarbonate bottles under semi-continuous conditions followed the protocol of Leong et al (2010) [17]. All cultures were maintained at 25 °C and at irradiance of 350  $\mu\text{mol m}^{-2} \text{s}^{-1}$  on a 12:12 hour light-dark cycle using cool white fluorescent tubes. Throughout the study, F/2 modified media [18] without silicate and containing varying concentrations of either nitrate or ammonium was used.

Prior to experimentation, cells were acclimated to modified media containing 3  $\mu\text{M-N}$  of nitrate and exposed to a light intensity of 350  $\mu\text{mol m}^{-2} \text{s}^{-1}$ . Acclimation took place over a period of 18 days with regular inoculation into fresh media every two days. *In vivo* chlorophyll-*a* levels were measured every two days to monitor growth rates.

Fluorescence readings were measured in triplicate using the Aquafluor™ fluorometer. All measurements were carried out in raw fluorescence units (RFU). Chlorophyll-*a* density

was subsequently derived by comparison against a calibration chart (Fig. 1). The linear relationship is represented by  $y = 2.94x + 3.356$  where  $y$  represents chlorophyll- $a$  density value and  $x$  represents RFU values.

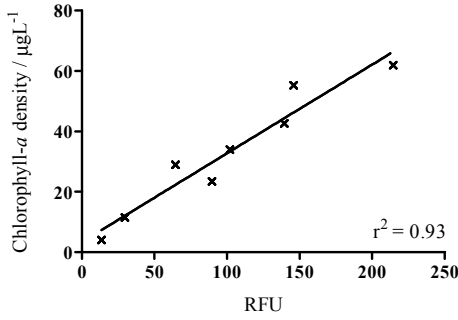


Fig. 1: Calibration chart to determine chlorophyll- $a$  density (y-axis) from RFU values (x-axis). RFU values were obtained by triplicate measurements using the Aquafluor™ fluorometer and chlorophyll- $a$  values are duplicate measurements following EPA Method 446.0 [19].

Light experiments were conducted by growing cells at a nitrate concentration level of  $3 \mu\text{M-N}$ . Low-light was simulated by using neutral light density filter which attenuated incubator light by about half. High-light conditions were achieved by directly exposing the culture bottles to the incubator light settings.

Nutrient experiments were conducted by inoculating exponentially-growing cultures into modified media containing  $3, 6$  and  $9 \mu\text{M-N}$  of nitrate or ammonium.

For each variable, three sets of cultures were grown over a period of five days. Specific growth rates ( $\mu \text{ day}^{-1}$ ) were calculated from fluorescence measurements taken at regular twenty-four intervals on each of the five days according to the equation:

$$\mu = \frac{\ln(N_t) - \ln(N_0)}{t_1 - t_0} \quad (1)$$

where  $N_t$  and  $N_0$  are the chlorophyll- $a$  density in  $\mu\text{g/L}$  on the day of measurement ( $t_1$ ) and at the start of the culture ( $t_0$ ). For each experimental culture, the initial chlorophyll- $a$  density was kept below  $6.3 \mu\text{g/L}$  ( $<10$  RFU).

### III. RESULTS

#### A. Ammonium vs. nitrate experiments

*Heterosigma* cells that were cultured under nitrate-enriched (Fig. 2) and ammonium-enriched (Fig. 3) conditions were observed to enter stationary growth phase by Day 3. There was no evident lag phase when nitrate-acclimated cells were inoculated into ammonium-enriched media. Exponential growth rates were found, using a two-way ANOVA test, to be significantly affected by both factors of nitrogen concentration levels ( $p < 0.0001$ ; 83.27% influence on total variation) and nitrogen type ( $p < 0.01$ ; 7.21% influence on total variation). There was insignificant interaction between the two factors. The values obtained are presented in Table 1.

TABLE I. EXPONENTIAL GROWTH RATES OF EXPERIMENTAL CULTURES

Conc. ( $\mu\text{M}$ )	$\text{NO}_3$ -enriched medium		$\text{NH}_4$ -enriched medium	
	Mean ( $\text{day}^{-1}$ )	S.D.	Mean ( $\text{day}^{-1}$ )	S.D.
3	0.70	0.1	0.66	0.04
6	1.08	0.06	0.89	0.1
9	1.25	0.07	1.10	0.05

Cultures inoculated in nitrate-enriched media yielded higher growth rates than ammonium-enriched media. Using Michaelis-Menton kinetics (Fig. 4), nitrate-enriched cultures attained maximum growth rate ( $V_{\text{max}}$ ) of  $2.06 \pm 0.2 \text{ day}^{-1}$  with half-saturation constant ( $K_m$ ) of  $5.71 \pm 0.9 \mu\text{M-NO}_3$ . Ammonium-enriched cultures attained a  $V_{\text{max}}$  of  $1.65 \pm 0.2 \text{ day}^{-1}$  with  $K_m$  of  $4.71 \pm 1.3 \mu\text{M-NH}_4$ .

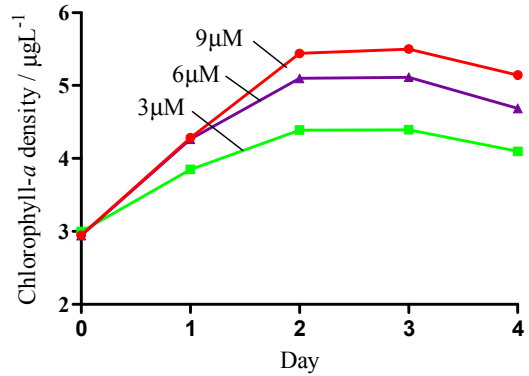


Fig. 2: Growth curves for *Heterosigma* at  $3 \mu\text{M-NO}_3$  (green closed squares),  $6 \mu\text{M-NO}_3$  (purple closed triangles) and  $9 \mu\text{M-NO}_3$  (red closed circles). Chlorophyll- $a$  density values were obtained by conversion of RFU readings using the calibration chart. Values are plotted against a logarithmic vertical axis. The points shown depict mean values.

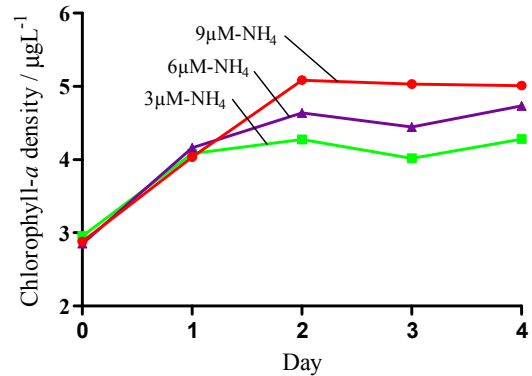


Fig. 3: Growth curves for *Heterosigma* at  $3 \mu\text{M-NH}_4$  (green closed squares),  $6 \mu\text{M-NH}_4$  (purple closed triangles) and  $9 \mu\text{M-NH}_4$  (red closed circles). Values are plotted against a logarithmic vertical axis. Points presented are derived from mean values.

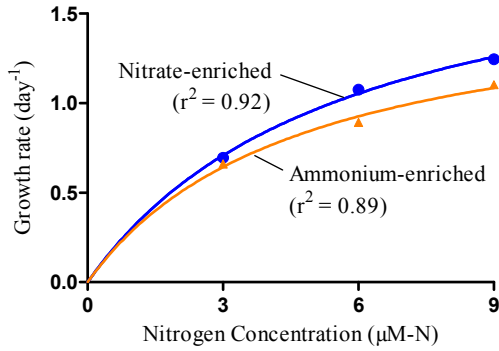


Fig. 4: Growth rates of *Heterosigma* in relation to increasing nitrate (blue closed circles) and ammonium (orange closed triangles) concentration levels. The line is derived according to Michaelis-Menten kinetics and points depict mean values.

### B. High-light vs. low-light experiments

Exponential growth rates were found to be statistically similar for cultures grown under low-light and high-light conditions. The obtained results are presented in Table 2.

TABLE II. EFFECT OF LIGHT ON EXPONENTIAL GROWTH RATES (DAY<sup>-1</sup>)

Light levels	Exponential growth (day <sup>-1</sup> )		
	Mean	S.D.	p-value
High	0.70	0.1	0.5
Low	0.76	0.1	

a. p-values are calculated using single-factor ANOVA

In general, the differences in the sampling parameters are statistically-insignificant, demonstrating the tolerance of *Heterosigma* cultures over the experimental range of light levels. Cultures grown under high-light (Fig. 5A) were also observed to have a sharper exit from the exponential growth phase. Under low-light conditions (Fig. 5B), cultures exhibited a more prolonged exponential growth phase.

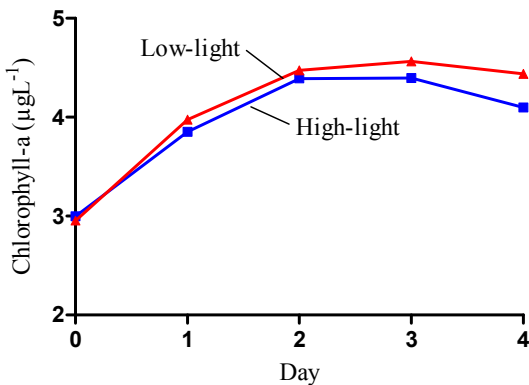


Fig. 5: Growth curves for *Heterosigma* under high-light (blue closed squares) and low-light (red closed triangles) conditions. Chlorophyll-a density values are plotted against a logarithmic vertical axis. Points depict mean values.

## Discussion

For the experiment between ammonium and nitrate sources of nitrogen, the present study has demonstrated that *Heterosigma* exhibited a preference for nitrate compared to ammonium. Nevertheless, both sources were suitable for the growth of *Heterosigma* cultures. Nitrate-uptake behaviour is understood to be thermodynamically more costly than ammonium uptake [20]. *Heterosigma* cells used in this study were therefore able to cope with the higher energy demand associated with nitrate-uptake. Although separate studies have indicated similar trends for several phytoplankton species [21], other studies have observed *Heterosigma* exhibiting a stronger preference for ammonium sources [22-23]. Such contrasting observations may indicate strain-specific biological mechanisms that may not yet be well-understood.

In addition, the absence of a lag phase when nitrate-acclimated *Heterosigma* cells were inoculated into ammonium-enriched media indicates its ability to rapidly switch from nitrate-uptake to ammonium-uptake behaviour. Conversely, inoculation of ammonium-acclimated cells into nitrate-enriched media has been found to require a lag phase associated with the physiological development of nitrate-uptake carrier proteins [24]. The uptake of nitrate and ammonium is therefore likely to require different physiological features.

In addition, this study also observed that *Heterosigma* cells cultured under low light intensity had growth rates that did not significantly differ from cells cultured under high light intensity, indicating the ability of *Heterosigma* cells to tolerate shaded conditions.

At present, the coastal waters of Singapore are dominated by non-toxic phytoplankton species [11]. The high diversity of local species assemblage may therefore serve to alleviate HAB-formation by decreasing the availability of open niches [25-27]. However, new niches can be created by human-driven changes to the coastal environment. For instance, low light conditions typically unfavourable to most marine algae [28] are being enhanced by sedimentation brought about through activities like reclamation, dredging and ship movement. The effect is compounded by N-pollution particularly through underground discharge which can provide suitable environments at depth for *Heterosigma* growth.

In addition, studies have also observed a relationship between industrialisation activities and increased nitrate pollution. Increased NO<sub>x</sub> emissions have led to nitrate-spiking of hydrological systems which discharge into coastal waters [e.g. 29-31]. If nitrate pollution is similarly present in Singapore, it could cause rapid growth of *Heterosigma* cells, increasing the opportunities for toxic algal bloom events.

The Singapore strains of *Heterosigma* are observed to be tolerant of a wide range of physical conditions and are hence able to rapidly colonise a wide range of niches. In order to ameliorate *Heterosigma* bloom events, it is important that favourable conditions are not provided and that open niches do not become available. Understanding the physiological responses of *Heterosigma* strains to various N sources as well as light intensity is an important step towards the mitigation of

bloom behaviour. Policy measures are therefore crucial to regulating the human-modification of coastal environments and to promoting the survival of the prevalent species assemblage.

### Acknowledgements

We thanked the staff and interns of TMSI; Ms SM Chew, Ms P Chen and Mr A Lee for assisting and overseeing the set-up and execution of the various experiments.

### Resources

- [1] Bowers, H. a., Tomas, C., Tengs, T., Kempton, J. W., Lewitus, A. J., & Oldach, D. W. (2006). Raphidophyceae [Chadefaud Ex Silva] Systematics and Rapid Identification: Sequence Analyses and Real-Time Pcr Assays. *Journal of phycology*, 42(6), 1333-1348. doi:10.1111/j.1529-8817.2006.00285.x
- [2] Rensel, J. E. J. (2007). *Fish kills from the harmful alga Heterosigma akashiwo in Puget Sound: Recent blooms and review*. Washington, USA. Retrieved from <http://www.whoi.edu/files/server.do?id=39383&pt=2&p=29109>
- [3] Kempton, J., Keppler, C. J., Lewitus, A., Shuler, A., & Wilde, S. (2008). A novel *Heterosigma akashiwo* (Raphidophyceae) bloom extending from a South Carolina bay to offshore waters. *Harmful Algae*, 7(2), 235-240. doi:10.1016/j.hal.2007.08.003
- [4] Black, E. A., White, J. N. C., Bagshaw, J. W., & Ginther, N. G. (1991). The effects of *Heterosigma akashiwo* on *Oncorhynchus tshawytscha* and its implications for fish culture. *Journal of Applied Ichthyology*, 7, 168-175.
- [5] Yang, C. Z., Albright, L. J., & Yousif, A. N. (1995). Oxygen-radical-mediated effects of the toxic phytoplankton *Heterosigma carterae* on juvenile rainbow trout *Oncorhynchus mykiss*. *Diseases of Aquatic Organisms*, 23, 101-108. doi:10.3354/dao023101
- [6] Khan, S., Arakawa, O., & Onoue, Y. (1997). Neurotoxins in a toxic red tide of *Heterosigma akashiwo* (Raphidophyceae) in Kagoshima Bay, Japan. *Aquaculture Research*, 28(1), 9-14. doi:10.1111/j.1365-2109.1997.tb01309.x
- [7] Twiner, M. J., Chidiac, P., Dixon, S. J., & Trick, C. G. (2005). Extracellular organic compounds from the ichthyotoxic red tide alga *Heterosigma akashiwo* elevate cytosolic calcium and induce apoptosis in Sf9 cells. *Harmful Algae*, 4(4), 789-800. doi:10.1016/j.hal.2004.12.006
- [8] The Ecology and Oceanography of Harmful Algal Blooms: Multidisciplinary Approaches to Research and Management by Donald M. Anderson. IOC Technical Series 74, UNESCO 2007. (English only). IOC/2007/TS/74
- [9] Masó, M., & Garcés, E. (2006). Harmful microalgae blooms (HAB); problematic and conditions that induce them. *Marine pollution bulletin*, 53(10-12), 620-630. doi:10.1016/j.marpolbul.2006.08.006
- [10] Tang, Y.Z. & Holmes, M.J. (2004). Cultured isolates of the raphidophytes *Chattonella marina*, *Fibrocapsa japonica* and *Heterosigma akashiwo* isolated from Singapore waters were non-toxic to fish. In Abstracts of the XI International Conference on Harmful Algal Blooms, Cape Town. pp. 244.
- [11] Gin, K. Y.-hoong, Holmes, M. J., Zhang, S., & Lin, X. (2006). Phytoplankton structure in the tropical port waters of Singapore. In E. Wolanski (Ed.), *The Environment in Asia Pacific Harbours* (pp. 347-375). Netherlands: Springer.
- [12] Lundholm, N., & Moestrup, Ø. (2006). The Biogeography of Harmful Algae. In E. Granéli & J. T. Turner (Eds.), *Ecology of Harmful Algae* (189th ed., Vol. 189, pp. 23-35). Heidelberg, Berlin: Springer. doi:10.1007/978-3-540-32210-8\_3
- [13] Ryther, J. H. & Dunstan, W. M. (1971). Nitrogen, Phosphorus, and Eutrophication in the Coastal Marine Environment. *Science*, 171, (3975), 1008-1013
- [14] Howarth, R. W. (2008). Coastal nitrogen pollution: A review of sources and trends globally and regionally. *Harmful Algae*, 8(1), 14-20. doi:10.1016/j.hal.2008.08.015
- [15] Case Study of Phytoplankton Blooms in Serangoon Harbor of Singapore by B. H. Ooi, H. Zheng, K. P. Yue, H. Kurniawati, P. Sundarambal, M. H. Dao, K. A. P. Roopsekhar, J. Wei, W. Cho, P. Tkalic, P. Malanotte-Rizzoli, and N. M. Patrikalakis. *OCEANS'10 IEEE Conference and Exhibition*, May 2010, Sydney, Australia.
- [16] Paerl, H. W. (1997). Coastal eutrophication and harmful deposition algal blooms: Importance of atmospheric deposition and groundwater as "new" nitrogen and other nutrient sources. *Limnology and Oceanography*, 42(5), 1154-1165.
- [17] Leong, S. C. Y., Maekawa, M., & Taguchi, S. (2010). Carbon and nitrogen acquisition by the toxic dinoflagellate *Alexandrium tamarense* in response to different nitrogen sources and supply modes. *Harmful Algae*, 9(1), 48-58. doi:10.1016/j.hal.2009.07.003
- [18] Guillard, R.R. & Ryther, J.H. (1962). Studies of marine planktonic diatoms. 1. *Cyclotella nana* Hustedt, and *Detonula confervacea* (Cleve) Gran. *Canadian Journal of Microbiology*, 8, 229-239.
- [19] Method 446.0 (Revision 1.2): In Vitro Determination of Chlorophylls a, b, c<sub>1</sub> + c<sub>2</sub> and Pheopigments in Marine and Freshwater Algae by Visible Spectrophotometry by Elizabeth J. Arar. U.S. Environmental Protection Agency (EPA) 1997. (pp. 1-26).
- [20] Guerrero, M. G., Vega, J. M., & Losada, M. (1981). The assimilatory nitrate-reducing system and its regulation. *Annual review of plant physiology*, 32, 169-204.
- [21] Dortch, Q. (1990). The interaction between ammonium and nitrate uptake in phytoplankton. *Marine Ecology Progress Series*, 61, 183-201.
- [22] Herndon, J., & Cochlan, W. P. (2007). Nitrogen utilization by the raphidophyte *Heterosigma akashiwo*: Growth and uptake kinetics in laboratory cultures. *Harmful Algae*, 6(2), 260-270. doi:10.1016/j.hal.2006.08.006
- [23] Wood, G. J., & Flynn, K. J. (1995). Growth of *Heterosigma carterae* (Raphidophyceae) on nitrate and ammonium at three photon flux densities: evidence for N stress in nitrate-growing cells. *Journal of phycology*, 31, 859-867.
- [24] Miyagi, N., Satoh, S., & Fujii, T. (1992). A Nitrate-Inducible Plasma Membrane Protein of a Marine Alga, *Heterosigma akashiwo*. *Plant Cell Physiology*, 33(7), 971-976.
- [25] Riegman, R., Boer, M. D., & Domis, L. D. S. (1996). Growth of harmful marine algae in multispecies cultures. *Journal of Plankton Research*, 18(10), 1851-1866.
- [26] Smayda, T. J., & Reynolds, C. S. (2001). Community assembly in marine phytoplankton: application of recent models to harmful dinoflagellate blooms. *Journal of Plankton Research*, 23(5), 447-461.
- [27] Grosholz, E. (2002). Ecological and evolutionary consequences of coastal invasions. *Trends in Ecology & Evolution*, 17(1), 22-27.
- [28] Margalef, R. (1978). Life-forms of phytoplankton as survival alternatives in an unstable environment. *Oceanologica acta.*, 134, 493-509.
- [29] Brimblecombe, P. & Stedman, D.H. (1982). Historical evidence for a dramatic increase in the nitrate component of acid rain. *Nature*, 298, 460-462. doi:10.1038/298460a0
- [30] Kaiser, J. (2001). The other global pollutant: Nitrogen proves tough to curb. *Science*, 294(5545), 1268-1269. doi: 10.1126/science.294.5545.1268
- [31] Morton, B. (1989). Pollution of the coastal waters of Hong Kong. *Marine Pollution Bulletin*, 20(7), 310-318.